



Deliverable of WP3

D.3.3: Review of models for predicting the concentration of chemicals in air, water and soil to human exposure, including mathematical and functional specification of the multimedia software

Contributors: Damiá Barceló, Daniel Guillén, Antoni Ginebreda, Rosa M. Darbra

Grant Agreement number: 226552

Project Acronym: RISKCYCLE

Project title: Risk-based management of chemicals and products in a circular economy at a global scale

Funding Scheme: 7th Framework Program (FP7)

Project starting date: 01 September 2009

Project duration: 36 months

Name of the scientific representative of the project's coordinator and organisation:

Prof. Dr.-Ing. habil. Dr. h.c. Bernd Bilitewski

Dresden University of Technology Institute of Waste Management and Contaminated Sites Treatment

Table of Contents

1. Introduction	4
2. Assessing environmental concentrations of chemicals: Measuring vs. Modelling	g 5
3. Use of exposure models in the risk assessment of chemicals	
4. Conclusions	
5. References	

1. Introduction

Our technological society makes extensive and intensive use of chemicals (most of them organics) and this number is continuously growing. Thus, for instance the European Inventory of Commercial Chemical Substances (EINECS) reports up today 100,204 commercially available substances [1] and similar figures hold for the U.S.A [2-3].

Hence, depending on their properties, mode and extent (volume) of use, this large amount of different chemicals can potentially reach the environment, having unpredictable environmental and health effects in long term. This has become a matter of major concern and constitutes the reason to release new regulations related to the safety of chemicals. Thus, for instance, the existing European Union regulation REACH (EC 1907/2006) [4] foresees to regulate chemicals used in commerce and consumer products, including a list of c.a. 30,000 compounds. About 10,000 have been already registered. From these, 2,782 are produced in large quantities (> 1,000 tons/year) [3].

On the other hand, a simultaneous and huge progress on the analytical methods and techniques has taken place, mostly associated to the development of multiresidue analytical methods based on chromatographic techniques (GC and LC) coupled to mass spectrometry (MS), capable to identify and quantify compounds at environmental trace levels of ng or pg/l. Such progress has substantially enlarged the possibilities of environmental monitoring and control. However, since not all measurable compounds are worth to be measured some kind of prioritisation or ranking is required in order to allocate analytical control efforts towards some target compounds, otherwise the task would be unbearable. The underlying rationale in the majority of the prioritisation lists of chemicals is based on the notion of risk assessment. Risk is broadly defined as the combination (i.e., product) of a probability of occurrence of some event by its hazard effects:

RISK = OCCURRENCE x EFFECTS

Correspondingly, the risk assessment process may be defined as the set of procedures aiming to identify hazards and to quantify the associated risk (in our

4

case, related to chemicals) concerning human health and/or ecosystems impairment (see Figure 1).



Figure 1. The risk assessment process

In the case of the environmental risk posed by chemicals, 'hazard effects' are related to the intrinsic properties of each compound [5] whereas 'occurrence' is associated to its environmental exposure, usually expressed in terms of environmental concentration.

Different risk assessment approaches have been developed in order to identify and rank compounds of environmental concern for both regulatory and monitoring purposes. Whereas most of all the existing schemes share the basic underlying risk assessment paradigm, they differ on how both factors, i.e., occurrence and effects, are defined and hence quantified.

2. Assessing environmental concentrations of chemicals: Measuring *vs.* Modelling

There are two basic approaches for establishing environmental concentrations, namely, measurement or modelling, the derived respective concentrations usually referred as MEC (Measured Environmental Concentrations) or the so called PEC (Predicted Environmental Concentrations). The most obvious and classical is

Deliverable 3.3: Review of models for predicting the concentration of chemicals in air, water and soil to human exposure, including mathematical and functional specification of the multimedia software through analytical chemistry obtaining MEC's. However, the development of environmental modelling provides an interesting alternative.

Essentially, the main drawback of measuring is that it is focussed on certain preselected analytes so that there is low chance of finding new targets. Other aspects of concern are their limited possibilities regarding time and space coverage, which can miss certain events. Sampling issues become thus important since analytical campaigns are usually expensive and time consuming.

Conversely, the main advantage of measurement is that it provides reliable results that are quite independent of the laboratories (at least, it is true for those that have an adequate Quality Assurance/Quality Control System). As far as modelling is concerned, it is fast and relatively affordable and has very good time and space coverage possibilities, being on the other hand its main weakness a strong dependence on the model chosen and the inherent uncertainty because of the lack of reliable data: physical-chemical properties, sources and fate.

Both options, measuring and modelling, have been discussed in detail by Johnson et al. [6] and their respective pros and cons are summarized in Table 1. However, since they are complementary, the wisest recommendation would be making use of both alternatives in order to exploit their respective advantages. Unfortunately, such desirable complementary approach is rarely seen together (see as instances, [7-8]).

6

Table 1: Comparative overview of strengths and weaknesses of analytical measurement vs. modelling in environmental studies (Adapted from reference [6]).

		-Results obtained reflect well reality.
		-Repeatability and reproducibility of results (at least between
		good qualified labs)".
		-Measurements are independent of information/data sources.
	Pros	-Multipurpose analytical methods can cover many compounds
		on a single run.
		-Even the best model will ultimately need to be experimentally
		checked.
		-Discovery of new emerging contaminants is possible.
Analytical		-Determination of compounds at very low quantities may be
measurement		difficult.
	Cons	-Time and space coverage require expensive monitoring
		campaigns.
		-Sampling campaigns may miss crucial episodes.
		-Analytical measurements give a snapshot, rather a continuous
		picture.
		-Expensive analytical equipment and method development.
		- Target monitoring may miss pollutants: "you only find what you
		are looking for"
		-Very good coverage capabilities of time and space.
		- Computation equipment is affordable.
		- Possibility of application to hypothetical scenarios: "What if?"
	Pros	- Useful for extrapolations to future (predictions on space and
Predicting		time, even for products not yet in the market).
(Modelling)		- Simultaneous modelling of many compounds.
		-Once the model is set up are fast and cheap to use.
		- Different models may render very different results.
	Cons	- Models are strongly dependent on parameter and data input.
		- Diffuse sources of pollution may be very difficult to model.

3. Use of exposure models in the risk assessment of chemicals

Current state of the art on spatially explicit multimedia fate models have been recently reviewed by Pistocchi et al. [9]. These authors distinguish three basic approaches:

- (i) Multiple box models.
- (ii) Numerical solutions to the advection-dispersion transport models.
- (iii) Meta-models. Geographic information system (GIS)-based modelling.

A full description and discussion of all the above models is beyond the scope of the present document (more information can be find in [9] and in other references cited therein). However, irrespectively of the type of model we are dealing with, they share some common requirements regarding to the input data. First of all, the basic environmental fate of a chemical compound is governed by its physical and chemical properties, such as partition constants, solubilities or kinetic constants characterizing the different dynamic phenomena like reactions (biodegradation, hydrolysis, photodegradation, etc.), adsorption, volatilization etc., as well as flow velocities, wind, temperature etc. On the other hand, the second major input information required is related to the physical factors associated to the receiving environment (i.e., temperature, humidity, wind speed etc.). Finally, a third group of factors is related the emission of the contaminant. Here, the amount of product released (consumption), the mode of use (closed cycle, spreading or chemical additives) and emission characteristics (point or diffuse) are very important. These aspects have been reviewed in a previous deliverable (Deliverable 3.2.: "Overview

8

of environmental factor influence over additive exposure and release into the environment")

The main weaknesses of most models rely on the great uncertainties embodied in many of those parameters. For this reason they are usually subjected to appropriate sensitivity analysis. Often, results obtained from modelling require further fine tuning of model parameters, which is only possible through empirical adjustment after experimental measurement. These models predict the distribution of a chemical between several environmental compartments and the final output of models will be the spatially distributed Predicted Environmental Concentration (PEC).

Exposure models are therefore valuable tools for indirect exposure assessment offering high versatility for quantifying risk associated to chemical exposure. Some of their already proved advantages are the following [10]:

- They allow predicting potential exposures for future or hypothetical releases/scenarios.
- They allow combining different types of contaminants and emission sources.
- The degree of complexity adopted by the model can be set according to the needs of the assessment.
- They consider exposures via multiple routes and pathways.
- They reduce the need for resource-intensive monitoring programmes.

As depicted in Figure 1, a typical risk assessment process involves the estimation of both the exposure (occurrence) and the hazard associated to the compound considered. According to the nature of each compound, they can exhibit several effects against the environment, including: persistence, bioaccumulation, ecotoxicity, endocrine disruption, mutagenicity, carcinogenicity, etc.

Correspondingly, modelling may be extended to both aspects and many of the most popular modelling existing software packages so do.

In Table 2 are listed the most widespread used models, together with their most positive characteristics and limitation. Further details can be found in the original references cited and in the document issued within (WP5), entitled "*Review of models used to assess human toxicity and ecotoxicological impacts of chemicals*" (Deliverable 5.2.).

Table 2: Strong and weak points of the risk assessment models gathered on the reviewed literature. (Some information adapted from Rosenbaum et al. who made a comparison between several models and the USEtox model [11]).

MODEL	STRENGTHS	WEAKNESS	REF	
Qwasi	 Equations available and easy to implement. Considers steady and unsteady states. 	- Human toxicity not considered and only considers one scale.	[12]	
Ecopoints	-	 Does not distinguish between categories (human, ecosystem). Does not realize fate and exposure analyses. 	[13],[14]	
ChemCAN	- Very transparent model.	- Only chemical fate model.	[15]	
ECOSENSE	 Most reliable modelling of classical air pollutants amongst the observed models. Bottom-up, i.e. spatially resolved, assessment capabilities for Europe, Russia, China/Asia, and Brazil/South America. 	 Does not consider mostly of organic chemicals and use an open system boundaries. Only inhalation exposures with respect to toxic impacts (additionally impacts on crops and building materials). 	[16]	
WMPT	- Key property based.	- Ecosystem toxicity not assessed. - No explicit fate results available and exposure routes not specified.	[17]	
EDIP	- Key property based and normalization and weighting methods provided.	Mainly representative for Europe.No explicit fate results available and no severity measure for human toxicity.	[18]	
Eco-	- Environmental problems defined at the endpoint level.	- It is assumed that all emissions and land uses, and their subsequent	[40]	
indicator 99	- Uncertainty analysis available.	damages occur in Europe.	[19]	
CSOIL 2000	 Simplicity of the model. Multiple human exposure ways considered. 	Ecotoxicity not considered.Fate and toxicological effects not considered.	[20]	
CalTOX	 Most encompassing in terms of exposure pathways. Advanced modelling of soil (several layers). Monte Carlo uncertainty estimation. 	 No severity measure for human toxicity, only partly compatible with damage approach- Ecosystem toxicity not assessed (e.g., marine environment and 	[21]	

		coastal zone not included for fate modelling).	
	- Continental average characterization factors available for		
IMPACT 2002+	different global regions		
	- Considering indoor air exposure	- Marine environment poorly represented so far.	[22]
	- Direct application of pesticides considered		
	- HC50 approach for effect modelling		
	- It is valuable for both live cycle risk assessment (LCA) and risk		
EUSES	assessment (RA).	- Difficulties in the operation of certain substances	[23]
	- There is a lot of media considered.		[23]
	- Uncertainty analysis available.		
	- Simplicity of the model.	- Ecotovicity not considered	
Humanex	- Multiple human exposure ways considered.	- Eate and toxicological effects not considered	[24]
Trantanex	- Capability on calculating maximum permissible concentrations	- Not utterly reliable when operating with pop-jonic organic chemicals	[24]
	(MPC) for the compounds of interest.		
XtraFOOD	- Age and gender categories are distinguished.	- Focused on the terrestrial food chain.	[25]
RAIDAR	- Simple to apply.	- Only most sensitive endpoints into consideration.	[2]
	- Capable of conducting full-chain risk assessments.		
2-FUN tool	- Pharmacokinetic models included.	- Ecosystem toxicity not assessed.	[26]
	- Probabilistic and sensitivity analyses considered.		
	- The user can choose the detail level of the results (midpoint or		
ReCiPe	endpoint).	- Different scales not considered.	[27]
	- Uncertainty analysis available.		
	- HC50 approach for effect modelling.		
USEtox	- Effect of intermittent rain events estimated.	- Only toxicity impacts considered.	[28]
	- Consensus model.		

USES-LCA	 Marine environment included. HC50 approach for effect modelling. One-dimensional uncertainty factors available. 	- Global coverage not spatially resolved.	[29]
GLOBOX	 Spatial differentiation (separate countries and oceans). Metal-specific processes in freshwater and marine environments handled. Dynamic calculations possible. 	 Regions distinguished are very different in size and are characterised by a wide variation in environmental parameters. The modelling of export and import of food for determining the intake by humans requires data and assumptions that may introduce additional uncertainty. 	[30]
MAFRAM	Simple to apply.Species from different taxa considered for the ecotoxicity value computation.	 Specifically developed for agricultural chemicals. Risk categories defined as crisp numbers. 	[31]

4. Use of Quantitative Structure-Activity Relationships (QSAR)

Since experimental assays for determining substances effects to the environment and living forms are expensive, time-consuming and require testing on animals, risk assessors and toxicologists are using models as a tool for estimate exposure effects of chemicals. QSAR (Quantitative Structure-Activity Relationship) provides a valuable tool for predicting these effects. In QSAR models the chemical structures are quantitatively correlated with their physico-chemical properties (melting point, water solubility, etc.), environmental fate (hydrolysis, biodegradation, etc.), ecotoxicity (acute and chronic toxicity) and other activities related to human health (carcinogenicity, mutagenicity, etc.). Several programs based on QSAR are available for many endpoints. Table 3 lists the main softwares used for risk assessment procedures.

In 1989 the specification SMILES [32] (simplified molecular input line entry specification) was developed. It consists on a nomenclature for describing molecular structures. SMILES notation has been widely used as an input for modelling since it is a fast and easy way of introducing molecular structures.

QSAR models play an important role as an environmentally oriented approach for regulatory assessment, especially under the new regulations. For instance, the REACH regulation (European Commission, 2006) strongly recommends the use of QSARs. Therefore QSAR procedures represent a very useful tool for risk assessment since experimental data is not always available. In this way, results can be provided without using animal testing.

Most of the QSAR models described in the literature are applied to organic substances [33-35], however in a lesser degree, they seem to be applied to inorganic (metals and organometallic compounds) [36-38], probably due to the lack

of a suitable tool for calculating descriptors for heavy atoms [36]. Further detailed information regarding to the use of QSAR can be found in the document issued within (WP4), entitled "Report on the review of bioassays and biosensors and (Q)SAR models as candidate for the intended use" (Deliverable 4.3).

To facilitate practical application of (Q)SAR approaches in regulatory contexts by governments and industry and to improve their regulatory acceptance, the OECD has started the development of various outcomes such as the principles for the validation of (Q)SAR models, guidance documents as well as the QSAR Toolbox. This item is intended to make (Q)SAR methods readily accessible, transparent, and less demanding in terms of computation costs.

Table 3: QSAR software available for predicting targeted endpoints from structureactivity relationships.

QSAR software		Endpoints	Developer
ACD	"Advanced Chemistry Development"	 Physico chemical properties (logP, solubilities, vapour pressure, etc.). Toxicology. Bioaccumulation and biomagnification factors. 	[39]
CAChe (COSMOtherm, OpenTox, etc.)	"Computer-Aided Chemistry"	 Physico chemical properties Toxicology 	[40]
CAESAR	"Computer Assisted Evaluation of industrial chemical Substances According to Regulation"	 Environmental fate: bioaccumulation and bioconcentration factors. carcinogenicity and mutagenicity Skin sensitisation. 	[41]
CASE (MultiCASE, MCASE, CASETOX and TOXALERT)	"Computer automated structure evaluation"	 Acute toxicity, carcinogenicity, teratogenicity, ecotoxicity, genetic toxicity. Biodegradation, bioaccumulation. Enzyme inhibition. Skin, eye irritations and allergies. 	[42-43]
COMPACT	"Computer-optimized molecular parametric analysis of chemical toxicity"	- Potential toxicity and carcinogenicity.	[44-45]
DEMETRA		- Ecotoxicity.	[46]
DEREK	"Deductive estimation of risk from existing knowledge"	 Toxicological, including: carcinogenicity, mutagenicity and teratogenicity. Skin sensitization, , irritancy, and respiratory sensitization 	[47]
DRAGON		 Physico chemical properties Toxicity. 	[48]

		- Fate: bioaccumulation, biodegradation rates,	
		partitioning among	
		environmental compartments, etc.	
EPI suite		- Physico chemical properties.	
(AOPWIN,	"Estimation Program	- I oxicity.	
BIOWIN,	Interface" suite	- Degradation rates (photolysis, biodegradation	[49]
HYDROWIN, etc)		and hydrolysis).	
		- Fate: bioaccumulation, biodegradation, etc.	
		- Toxicity, also estimates	
HazardExpert		- Toxicokinetic effects: bioaccumulation and	[50]
		bioavailability on	[]
		the basis of predicted physicochemical values.	
	"Optimized Approach	- Physicochemical properties.	
OASIS	Based on Structural	 Toxic endpoints accounting for conformational 	[51]
	Indices Set"	flexibility of Chemicals.	
OFCD QSAR		- (eco)toxicity	
Toolbox		- Skin sensitisation	[52]
		- Mutagenicity	
		- Chemical reactivity.	
OncoLogic		- Metabolic activation.	[54]
		- Mechanisms of chemical carcinogenesis [53].	
	"Substructure-Based	- Acute and chronic toxicity	
SUCCSES	Computerized	- Mutagenic and carcinogenic developmental	[55]
5000525	Chemical Selection	reproductive or neurotoxic effects	[00]
	Expert System"		
		- Psysico-chemical properties (logP).	
	"Toxicity Prediction by Komputer Assisted	- Environmental fate	
ТОРКАТ		- Acute and chronic toxicity, ecotoxicity,	[56]
	Technology"	- Carcinogenicity, mutagenicity, and	[]
	roomiology	reproductive/developmental effects.	
		- Skin and eye irritation .	

5. Conclusions

Model development for predicting the environmental occurrence, fate and effects caused by chemicals has been a continuously growing discipline during the last years. Models have expanded their domain of application beyond research and have themselves revealed as powerful tools in management, decision support and regulation development.

Current limitations on the practical use of fate models are mostly due to lack of information on chemical emissions [9]. Therefore this is one of the R&D needs identified within the RISKCYCLE project. Its relevance is still more acute, if one takes into consideration the scenario of circulation of goods and products (and Deliverable 3.3: Review of models for predicting the concentration of chemicals in air, water and soil to human exposure, including mathematical and functional specification of the multimedia software consequently of chemicals) at earth scale through either natural or anthropogenic processes which has become the most significant characteristics of our global world.

Acknowledgements

The authors would like to acknowledge the information gathered by other Riskcycle partners that has been summarized in the present document.

6. References

- 1. European Inventory of Existing Commercial Chemical Substances of 15 June 1990. OJ C/146A. Corrigendum published in OJ C54, 2002.
- 2. Arnot, J.A., et al., *Screening Level Risk Assessment Model for Chemical Fate and Effects in the Environment*. Environmental Science & Technology, 2006. **40**(7): p. 2316-2323.
- 3. Muir, D.C.G. and P.H. Howard, Are There Other Persistent Organic Pollutants? A Challenge for Environmental Chemists[†]. Environmental Science & Technology, 2006. **40**(23): p. 7157-7166.
- 4. EU, Regulation 1907/2006 of the European Parliament and of the council of 18 December 2006 concerning the Registration, Evaluation, Authorisation and Restriction of Chemicals (REACH), establishing a European Chemicals Agency, amending Directive 1999/45/EC and repealing Council Regulation (EEC) No 793/93 and Commission Regulation (EC) No 1488/94 as well as Council Directive 76/769/EEC and Commission Directives 91/155/EEC, 93/67/EEC, 93/105/EC and 2000/21/EC. 2006.
- 5. Arnot, J.A. and D. Mackay, *Policies for Chemical Hazard and Risk Priority Setting: Can Persistence, Bioaccumulation, Toxicity, and Quantity Information Be Combined?* Environmental Science & Technology, 2008. **42**(13): p. 4648-4654.
- 6. Johnson, A.C., et al., Assessing the Concentrations of Polar Organic Microcontaminants from Point Sources in the Aquatic Environment: Measure or Model? Environmental Science & Technology, 2008. **42**(15): p. 5390-5399.
- 7. Coetsier, C.M., et al., *Discharge of pharmaceutical products (PPs) through a conventional biological sewage treatment plant: MECs vs PECs?* Environment International, 2009. **35**(5): p. 787-792.
- 8. Bound, J.P. and N. Voulvoulis, *Predicted and measured concentrations for selected pharmaceuticals in UK rivers: Implications for risk assessment*. Water Research, 2006. **40**(15): p. 2885-2892.
- 9. Pistocchi, A., D.A. Sarigiannis, and P. Vizcaino, *Spatially explicit multimedia fate models for pollutants in Europe: State of the art and perspectives.* Science of The Total Environment, 2010. **408**(18): p. 3817-3830.
- 10. Fryer, M., et al., *Human exposure modelling for chemical risk assessment: a review of current approaches and research and policy implications.* Environmental Science & Policy, 2006. **9**(3): p. 261-274.
- Rosenbaum, R., et al., USEtox—the UNEP-SETAC toxicity model: recommended characterisation factors for human toxicity and freshwater ecotoxicity in life cycle impact assessment. The International Journal of Life Cycle Assessment, 2008. 13(7): p. 532-546.
- 12. Mackay, D., M. Joy, and S. Paterson, *A quantitative water, air, sediment interaction* (*QWASI*) fugacity model for describing the fate of chemicals in lakes. Chemosphere, 1983. **12**(7-8): p. 981-997.
- 13. Ahbe, S., Methodik für Ökobilanzen auf der Basis ökologischer Optimierung. 1990.
- 14. Braunschweig, A.M.-W.R., Ökobilanzen für Unternehmungen: eine Wegleitung für die Praxis, ed. P.H. Verlag. 1993.
- Mackay, D., et al., Assessment of chemical fate in the environment using evaluative, regional and local-scale models: Illustrative application to chlorobenzene and linear alkylbenzene sulfonates. Environmental Toxicology and Chemistry, 1996. 15(9): p. 1638-1648.

- 16. Krewitt, W., A. Trukenmueller, P. Mayerhofer, et al. . H.Kremers & W.Pillmann, , *ECOSENSE—an integrated tool for environmental impact analysis. In Space and Time in Environmental Information Systems.* Umwelt-Informatik aktuell, Band 7. Metropolis-Verlag. Marburg, 1995.
- 17. Documentation., U.E.U.s.G.a.S., *Waste Minimization Prioritization Tool (Beta Version 1.0)*. EPA530-R-97-019, 1997.
- 18. Hauschild, M.a.H.W., *Environmental Assessment of products. Volume 2:* . Scientificbackground. Chapman & Hall, London, 1998.
- Hofstetter, P., Perspective in life cycle impact assessment. A structured approach to combine models of the technosphere, ecosphere and valuesphere. PhD Thesis, 1998.
 ETH Zurich (Swiss Federal Institute of Technology), Kluwer Academic Publishers, Dordrecht.
- 20. Brandes, L., H.d. Hollander, and D.v.d. Meent, *SimpleBox 2.0: a nested multimedia fate model for evaluating the environmental fate of chemicals.*
- 21. McKone, T., CalTOX, a multimedia total exposure model for hazardous-waste sites, Part I: executive summary. A report written for The Office of Scientific Affairs Depart-ment of Toxic Substances Control California Environmental Protection Agency Sacra-mento, California by the Lawrence Livermore National Laboratory., 1993a.
- 22. Jolliet O, F.G., Frischknecht R, Giegrich J, Guinée JB, Hauschild M, Heijungs R, Hofstetter P, Jesen AA, Lindeijer E, Müller-Wenk R, Nichols Ph, Potting J, Wenzel C, White P., *Impact assessment of human and eco-toxicity in Life Cycle Assessment. In: Towards a methodology for Life Cycle Impact Assessment.* SETAC, Brussels, 1996.
- 23. RIVM, V., WVC., Uniform System for the Evaluation of Substances (USES), version 1.0. National Institute of Public Health and the Environment (RIVM), Ministry of Housing, Physical Planning and Environment (VROM), Ministry of Welfare, Health and Cultural Affairs (WVC). The Hague, Ministry of Housing, Physical Planning and Environment. Distribution No. 11144/150 (1994).
- 24. Bontje, T.P.T.a.W.M., A human exposure model to calculate harmonized risk limits. Model description and analysis. . RIVM report 601501022, 2005.
- 25. Seuntjens P, S.W., Vangronsveld J., *Chain model for the impact analysis of contaminants in primary food products.* Study report of the Belgian Science Policy, 2006.
- 26. EU, European Union 6th Framework Programme. 2-FUN project (Full-chain and UNcertainty Approaches for Assessing Health Risks in FUture ENvi-ronmental Scenarios: <u>www.2-fun.org</u>). Contract n° FP6-2005-GLOBAL-4-036976, Feb 2007 Jan 2011.
- 27. Goedkoop, H., Huijbregts, Schryver, Struijs, Zelm. , *ReCiPe 2008: A life cycle impact assessment method which comprises harmonised category indicators at the mid-point and the endpoint level. First edition, Report I: Characterisation.* 2009.
- 28. UNEP-SETAC, United Nations Environment Pro-gram (UNEP) and Society for Environmental Toxicology and Chemistry (SETAC) consensus. USEtox Life Cycle Initiative. 2005.
- 29. Huijbregts, M.A.J., et al., Priority assessment of toxic substances in life cycle assessment. Part I: Calculation of toxicity potentials for 181 substances with the nested multi-media fate, exposure and effects model USES-LCA. Chemosphere, 2000. **41**(4): p. 541-573.
- 30. Wegener Sleeswijk, A. and R. Heijungs, *GLOBOX: A spatially differentiated global fate, intake and effect model for toxicity assessment in LCA*. Science of The Total Environment, 2010. **408**(14): p. 2817-2832.

- 31. Batiha, M.A., et al., *MAFRAM--A new fate and risk assessment methodology for non-volatile organic chemicals*. Journal of Hazardous Materials, 2010. **181**(1-3): p. 1080-1087.
- 32. Weininger, D., A. Weininger, and J.L. Weininger, *SMILES. 2. Algorithm for generation of unique SMILES notation*. Journal of Chemical Information and Computer Sciences, 1989. **29**(2): p. 97-101.
- 33. Dashtbozorgi, Z. and H. Golmohammadi, *Prediction of air to liver partition coefficient for volatile organic compounds using QSAR approaches*. European Journal of Medicinal Chemistry, 2010. **45**(6): p. 2182-2190.
- 34. Qin, H., et al., *Development and assessment of quantitative structure-activity relationship models for bioconcentration factors of organic pollutants*. Chinese Science Bulletin, 2009. **54**(4): p. 628-634.
- 35. Kusic, H., et al., *Prediction of rate constants for radical degradation of aromatic pollutants in water matrix: A QSAR study.* Chemosphere, 2009. **75**(8): p. 1128-1134.
- 36. Toropova, A.P., et al., *Co-evolutions of correlations for QSAR of toxicity of organometallic and inorganic substances: An unexpected good prediction based on a model that seems untrustworthy.* Chemometrics and Intelligent Laboratory Systems, 2011. **105**(2): p. 215-219.
- 37. Toropov, A., A. Toropova, and E. Benfenati, *QSAR-modeling of toxicity of organometallic compounds by means of the balance of correlations for InChI-based optimal descriptors*. Molecular Diversity, 2010. **14**(1): p. 183-192.
- Lepădatu, C., M. Enache, and J.D. Walker, *Toward a More Realistic QSAR Approach to Predicting Metal Toxicity*. QSAR & Combinatorial Science, 2009. 28(5): p. 520-525.
- 39. ACD, V.10 available from <u>http://www.acdlabs.com/</u>.
- 40. CAChe Research. Computer-Aided Chemistry & Biochemistry. CAChe software available at: <u>http://www.cacheresearch.com/sw.html</u>.
- 41. CAESAR EC project. Available at <u>http://www.caesar-project.eu</u>.
- 42. Klopman, G., *MULTICASE 1. A Hierarchical Computer Automated Structure Evaluation Program.* Quantitative Structure-Activity Relationships, 1992. **11**(2): p. 176-184.
- 43. Klopman, G. and H. Rosenkranz, *Structure-activity relations: maximizing the usefulness of mutagenicity and carcinogenicity databases*. Environmental health perspectives, 1991. **96**: p. 67-75.
- 44. Lewis, D.F.V., C. Ioannides, and D.V. Parke, *An Improved and Updated Version of the Compact Procedure for the Evaluation of P450-Mediated Chemical Activation*. Drug Metabolism Reviews, 1998. **30**(4): p. 709-737.
- 45. Lewis, D.F.V., *COMPACT: a structural approach to the modelling of cytochromes P450 and their interactions with xenobiotics.* Journal of Chemical Technology & Biotechnology, 2001. **76**(3): p. 237-244.
- 46. *DEMETRA EC project. Available at <u>http://www.demetra-tox.net.</u>*
- 47. Derek Nexus available from https://www.lhasalimited.org/.
- 48. DRAGON software. Virtual Computational Chemistry Laboratory. Available at <u>http://www.vcclab.org/lab/edragon/</u>.
- 49. U.S EPA. Estimation Program Interface (EPI) Suite. Available at <u>http://www.epa.gov/opptintr/exposure/pubs/episuite.htm</u>.
- 50. *Hazardexpert. Compudrug software corporation. software available at* <u>http://www.compudrug.com/?q=node/90</u>.
- 51. Oasis Basic Laboratory of Mathematical Chemistry Software available at: <u>http://oasis-lmc.org/?section=software&swid=12</u>.

- 52. The OECD QSAR Toolbox. Available at <u>http://www.oecd.org/document/54/0,3746,en_2649_37465_42923638_1_1_1_37465</u>,00.html.
- 53. Patlewicz, G., R. Rodford, and J.D. Walker, *Quantitative structure-activity relationships for predicting mutagenicity and carcinogenicity*. Environmental Toxicology and Chemistry, 2003. **22**(8): p. 1885-1893.
- 54. OncoLogic. Available at: <u>http://www.epa.gov/oppt/sf/pubs/oncologic.htm</u>.
- 55. J.D.Walker and D.A.Gray, *QSARs for Predicting Biological Activities Related to Endocrine Disruption and Environmental Human Health Interactions.* SETAC Press, 2002. **Pensacola, FL**.
- 56. TOPKAT software. Accelrys company. Software available at: <u>http://accelrys.com/products/discovery-studio/predictive-toxicology.html</u>.